

UNDERSTANDING VARIATION IN PARTITION COEFFICIENT, K_d, VALUES

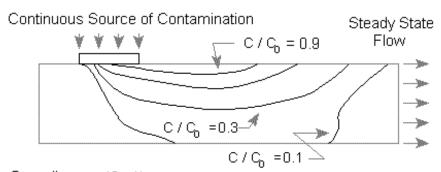


Volume I:

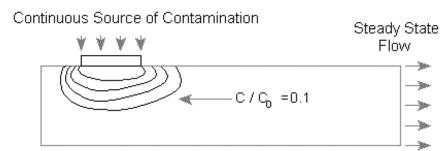
The \mathbf{K}_{d} Model, Methods of Measurement, and

Application of Chemical Reaction Codes

Case I: K_d = 1 ml/g



Case II: K_d = 10 ml/g



UNDERSTANDING VARIATION IN PARTITION COEFFICIENT, K_d, VALUES

Volume I:

The K_d Model, Methods of Measurement, and Application of Chemical Reaction Codes

August 1999

A Cooperative Effort By:

Office of Radiation and Indoor Air
Office of Solid Waste and Emergency Response
U.S. Environmental Protection Agency
Washington, DC 20460

Office of Environmental Restoration U.S. Department of Energy Washington, DC 20585

NOTICE

The following two-volume report is intended solely as guidance to EPA and other environmental professionals. This document does not constitute rulemaking by the Agency, and cannot be relied on to create a substantive or procedural right enforceable by any party in litigation with the United States. EPA may take action that is at variance with the information, policies, and procedures in this document and may change them at any time without public notice.

Reference herein to any specific commercial products, process, or service by trade name, trademark, manufacturer, or otherwise, does not necessarily constitute or imply its endorsement, recommendation, or favoring by the United States Government.

FOREWORD

Understanding the long-term behavior of contaminants in the subsurface is becoming increasingly more important as the nation addresses groundwater contamination. Groundwater contamination is a national concern as about 50 percent of the United States population receives its drinking water from groundwater. It is the goal of the Environmental Protection Agency (EPA) to prevent adverse effects to human health and the environment and to protect the environmental integrity of the nation's groundwater.

Once groundwater is contaminated, it is important to understand how the contaminant moves in the subsurface environment. Proper understanding of the contaminant fate and transport is necessary in order to characterize the risks associated with the contamination and to develop, when necessary, emergency or remedial action plans. The parameter known as the partition (or distribution) coefficient (K_d) is one of the most important parameters used in estimating the migration potential of contaminants present in aqueous solutions in contact with surface, subsurface and suspended solids.

This two-volume report describes: (1) the conceptualization, measurement, and use of the partition coefficient parameter; and (2) the geochemical aqueous solution and sorbent properties that are most important in controlling adsorption/retardation behavior of selected contaminants. Volume I of this document focuses on providing EPA and other environmental remediation professionals with a reasoned and documented discussion of the major issues related to the selection and measurement of the partition coefficient for a select group of contaminants. The selected contaminants investigated in this two-volume document include: chromium, cadmium, cesium, lead, plutonium, radon, strontium, thorium, tritium (3H), and uranium. This two-volume report also addresses a void that has existed on this subject in both this Agency and in the user community.

It is important to note that soil scientists and geochemists knowledgeable of sorption processes in natural environments have long known that generic or default partition coefficient values found in the literature can result in significant errors when used to predict the absolute impacts of contaminant migration or site-remediation options. Accordingly, one of the major recommendations of this report is that for site-specific calculations, <u>partition coefficient values measured at site-specific conditions are absolutely essential</u>.

For those cases when the partition coefficient parameter is not or cannot be measured, Volume II of this document: (1) provides a "thumb-nail sketch" of the key geochemical processes affecting the sorption of the selected contaminants; (2) provides references to related key experimental and review articles for further reading; (3) identifies the important aqueous- and solid-phase parameters controlling the sorption of these contaminants in the subsurface environment under oxidizing conditions; and (4) identifies, when possible, minimum and maximum conservative partition coefficient values for each contaminant as a function of the key geochemical processes affecting their sorption.

This publication is the result of a cooperative effort between the EPA Office of Radiation
and Indoor Air, Office of Solid Waste and Emergency Response, and the Department of Energy
Office of Environmental Restoration (EM-40). In addition, this publication is produced as part of
ORIA's long-term strategic plan to assist in the remediation of contaminated sites. It is published
and made available to assist all environmental remediation professionals in the cleanup of
groundwater sources all over the United States.

Stephen D. Page, Director Office of Radiation and Indoor Air

ACKNOWLEDGMENTS

Ronald G. Wilhelm from ORIA's Center for Remediation Technology and Tools was the project lead and EPA Project Officer for this two-volume report. Paul Beam, Environmental Restoration Program (EM-40), was the project lead and sponsor for the Department of Energy (DOE). Project support was provided by both DOE/EM-40 and EPA's Office of Remedial and Emergency Response (OERR).

EPA/ORIA wishes to thank the following people for their assistance and technical review comments on various drafts of this report:

Patrick V. Brady, U.S. DOE, Sandia National Laboratories

David S. Brown, U.S. EPA, National Exposure Research Laboratory

Joe Eidelberg, U.S. EPA, Region 9

Amy Gamerdinger, Washington State University

Richard Graham, U.S. EPA, Region 8

John Griggs, U.S. EPA, National Air and Radiation Environmental Laboratory

David M. Kargbo, U.S. EPA, Region 3

Ralph Ludwig, U.S. EPA, National Risk Management Research Laboratory

Irma McKnight, U.S. EPA, Office of Radiation and Indoor Air

William N. O'Steen, U.S. EPA, Region 4

David J. Reisman, U.S. EPA, National Risk Management Research Laboratory

Kyle Rogers, U.S. EPA, Region 5

Joe R. Williams, U.S. EPA, National Risk Management Research Laboratory

OSWER Regional Groundwater Forum Members

In addition, special acknowledgment goes to Carey A. Johnston from ORIA's Center for Remediation Technology and Tools for his contributions in the development, production, and review of this document.

Principal authorship in production of this guide was provided by the Department of Energy's Pacific Northwest National Laboratory (PNNL) under the Interagency Agreement Number DW89937220-01-03. Lynnette Downing served as the Department of Energy's Project Officer for this Interagency Agreement. PNNL authors involved in this project include:

Kenneth M. Krupka Daniel I. Kaplan Gene Whelan R. Jeffrey Serne Shas V. Mattigod

TO COMMENT ON THIS GUIDE OR PROVIDE INFORMATION FOR FUTURE UPDATES:

Send all comments/updates to:

U.S. Environmental Protection Agency Office of Radiation and Indoor Air Attention: Understanding Variation in Partition (K_d) Values 401 M Street, SW (6602J) Washington, DC 20460

or

webmaster.oria@epa.gov

ABSTRACT

This two-volume report describes the conceptualization, measurement, and use of the partition (or distribution) coefficient, K_d, parameter, and the geochemical aqueous solution and sorbent properties that are most important in controlling adsorption/retardation behavior of selected contaminants. The report is provided for technical staff from EPA and other organizations who are responsible for prioritizing site remediation and waste management decisions. Volume I discusses the technical issues associated with the measurement of K_d values and its use in formulating the retardation factor, R_f . The K_d concept and methods for measurement of K_d values are discussed in detail in Volume I. Particular attention is directed at providing an understanding of: (1) the use of K_d values in formulating R_f , (2) the difference between the original thermodynamic K_d parameter derived from ion-exchange literature and its "empiricized" use in contaminant transport codes, and (3) the explicit and implicit assumptions underlying the use of the K_d parameter in contaminant transport codes. A conceptual overview of chemical reaction models and their use in addressing technical defensibility issues associated with data from K_d studies is presented. The capabilities of EPA's geochemical reaction model MINTEQA2 and its different conceptual adsorption models are also reviewed. Volume II provides a "thumb-nail sketch" of the key geochemical processes affecting the sorption of selected inorganic contaminants, and a summary of K_d values given in the literature for these contaminants under oxidizing conditions. The contaminants chosen for the first phase of this project include chromium, cadmium, cesium, lead, plutonium, radon, strontium, thorium, tritium (³H), and uranium. Important aqueous speciation, (co)precipitation/dissolution, and adsorption reactions are discussed for each contaminant. References to related key experimental and review articles for further reading are also listed.

CONTENTS

<u>Page</u>
NOTICE ii COREWORD iii ACKNOWLEDGMENTS v UTURE UPDATES vi ABSTRACT vii AIST OF FIGURES xii
IST OF TABLES xiv
.0 Introduction
.0 The K _d Model And Its Use In Contaminant Transport Modeling
2.1 Introduction
2.2 Aqueous Geochemical Processes 2.3 2.2.1 Aqueous Complexation 2.3 2.2.2 Oxidation-Reduction (Redox) Chemistry 2.5 2.2.3 Sorption 2.8 2.2.3.1 Adsorption 2.10 2.2.3.1.1 Ion Exchange 2.13 2.2.3.2 Precipitation 2.13
$ \begin{array}{cccccccccccccccccccccccccccccccccccc$
2.4 Effects of Unsaturated Conditions on Transport
2.5 Effects of Chemical Heterogeneity on Transport 2.33 2.5.1 Coupled Hydraulic and Chemical Heterogeneity 2.34
2.6 Diffusion
2.7 Subsurface Mobile Colloids2.372.7.1 Concept of 3-Phase Solute Transport2.372.7.2 Sources of Groundwater Mobile Colloids2.382.7.3 Case Studies of Mobile-Colloid Enhanced Transport of

		Metals and Radionuclides	2.39
	2.8	Anion Exclusion	2.39
	2.9	Summary	2.40
3.0	Me	ethods, Issues, and Criteria for Measuring K _d Values	3.1
	3.1	Introduction	3.1
	3.2	Methods for Determining K _d Values 3.2.1 Laboratory Batch Method 3.2.2 In-situ Batch Method 3.2.3 Laboratory Flow-Through Method 3.2.4 Field Modeling Method 3.2.5 K _{oc} Method	3.3 3.8 3.9 3.14
	3.3	Issues Regarding Measuring and Selecting K _d Values 3.3.1 Using Simple Versus Complex Systems to Measure K _d Values 3.3.2 Field Variability 3.3.3 The "Gravel Issue" 3.3.4 The "Colloid Issue" 3.3.5 Particle Concentration Effect	3.16 3.18 3.19 3.21
	3.4	Methods of Acquiring K _d Values from the Literature for Screening Calculations 3.4.1 K _d Look-Up Table Approach: Issues Regarding Selection of K _d Values from the Literature 3.4.2 Parametric K _d Approach 3.4.3 Mechanistic Adsorption Models	3.23 3.26
	3.5	Summary	3.28
4.0		roundwater Calibration Assessment Based on Partition Coefficients: erivation and Examples	4.1
	4.1	Introduction	4.1
	4.2	Calibration: Location, Arrival Time, and Concentration	4.1
	4.3	Illustrative Calculations to Help Quantify K _d Using Analytical Models	4.4

	4.3.4 Dispersion and the Partition Coefficient	4.10
	4.4 Modeling Sensitivities to Variations in the Partition Coefficient	4.11
	4.4.2 Partition Coefficient as a Calibration Parameter in Transport Modeling	4.12
	4.5 Summary	4.14
5.0	O Application of Chemical Reaction Codes	. 5.1
	5.1. Background	. 5.1
	5.1.1 Definition of Chemical Reaction Modeling	
	5.1.2 Reviews of Chemical Reaction Models	
	5.1.3 Aqueous Speciation-Solubility Versus Reaction Path Codes	. 5.4
	5.1.4 Adsorption Models in Chemical Reaction Codes	
	5.1.5 Output from Chemical Reaction Modeling	
	5.1.6 Assumptions and Data Needs	. 5.9
	5.1.7 Symposiums on Chemical Reaction Modeling	5.10
	5.2 MINTEQA2 Chemical Reaction Code	5 11
	5.2.1 Background	
	5.2.2 Code Availability	
	5.2.3 Aqueous Speciation Submodel	
	5.2.3.1 Example of Modeling Study	
	5.2.3.2 Application to Evaluation of K _d Values	
	5.2.4 Solubility Submodel	
	5.2.4.1 Example of Modeling Study	
	5.2.4.2 Application to Evaluation of K _d Values	
	5.2.5 Precipitation/Dissolution Submodel	
	5.2.5.1 Example of Modeling Study	
	5.2.5.2 Application to Evaluation of K _d Values	
	5.2.6 Adsorption Submodel	
	5.2.6.1 Examples of Modeling Studies	5.22
	5.2.6.2 Application to Evaluation of K _d Values	
	5.2.7 MINTEQA2 Databases	
	5.2.7.1 Thermodynamic Database	5.24
	5.2.7.1.1 Basic Equations	5.25
	5.2.7.1.2 Structure of Thermodynamic Database Files	
	5.2.7.1.3 Database Components	
	5.2.7.1.4 Status Relative to Project Scope	
	5.2.7.1.5 Issues Related to Database Modifications	
	5 2 7 2 Sorntion Database	5 33

5.2.7.2.1 Status Relative to Project Scope	5.33
5.2.7.2.2 Published Database Sources	
5.3 Adsorption Model Options in MINTEQA2	5 35
5.3.1 Electrostatic Versus Non-Electrostatic Models	
5.3.2 Activity Partition Coefficient (K _d) Model	
5.3.3 Activity Langmuir Model	
5.3.4 Activity Freundlich Model	
5.3.5 Ion Exchange Model	
5.3.6 Diffuse Layer Model	
5.3.7 Constant Capacitance Model	
5.3.8 Triple Layer Model	
5.4 Summary	5.50
6.0 References	. 6.1
Appendix A - Acronyms, Abbreviations, Symbols, and Notation	. A.1
Appendix B - Definitions	. B.1
Appendix C - Standard Method Used at Pacific Northwest National Laboratory for Measuring Laboratory Batch K. Values	. C.1

LIST OF FIGURES

	<u>Page</u>
Figure 2.1.	Diffuse double layer and surface charge of a mineral surface 2.11
Figure 2.2.	Four types of adsorption isotherm curves shown schematically in parlance of Giles <i>et al.</i> (1973)
Figure 2.3.	Schematic diagram for conceptual model of water distribution in saturated (top two figures) and unsaturated soils (bottom two figures) suggesting differences in the unsaturated flow regime (indicated by arrows) for soils with varying texture
Figure 2.4.	Development of hydraulic heterogeneity (decreasing φ_m) in unsaturated, non-aggregated soils with decreasing moisture saturation
Figure 3.1.	Procedure for measuring a batch K_d value (EPA 1991)
Figure 3.2.	Demonstration calculation showing affect on overall K_d by multiple species that have different individual K_d values and are kinetically slow at interconverting between each composition state
Figure 3.3.	Procedure for measuring a column K_d value
Figure 3.4.	Schematic diagram showing the relative concentrations of a constituent at the input source (figures on left) and in the effluent (figures on right) as a function of time for a pulse versus step input
Figure 4.1.	Relative relationships between input-data quality, output uncertainty, and types of problems addressed by each level of assessment 4.2
Figure 4.2.	Example illustrating a MEPAS 90 Sr calibration with K_d equaling 0.8 ml/g and $1 \text{ monitored-data point} \dots 4.13$
Figure 4.3.	Example illustrating MEPAS 90 Sr calibrations with K_d equaling 0.4 and 0.8 ml/g and several monitored-data points 4.14
Figure 5.1.	Distribution of dominant U(VI) aqueous species for leachates buffered at pH 7.0 by local ground water (Figure 5.1a) and at pH 12.5 by cement pore fluids (Figure 5.1b)
Figure 5.2.	Saturation Indices calculated for rutherfordine (UO ₂ CO ₂) as a function of

	pH for solution analyses from Sergeyeva et al. (1972)	5.17
Figure 5.3.	Maximum concentration limits calculated for total dissolved uranium as a function of pH based on the equilibrium solubilities of schoepite and uranophane	5.20
Figure 5.4.	Schematic representation of the triple layer model showing surface species and surface charge-potential relationships	5.37
Figure 5.5.	Schematic representation of the constant capacitance layer model showing surface species and surface charge-potential relationships	5.38

LIST OF TABLES

	<u>Page</u>
Table 2.1.	List of several redox-sensitive metals and their possible valence states in soil/groundwater systems
Table 2.2.	Sequence of Principal Electron Acceptors in neutral pH aquatic systems (Sposito 1989)
Table 2.3.	Zero-point-of-charge, pH _{zpc}
Table 2.4.	Cation exchange capacities (CEC) for several clay minerals (Grim 1968) 2.14
Table 2.5.	Summary of chemical processes affecting attenuation and mobility of contaminants
Table 3.1.	Representative chemical species in acidic and basic soil solutions (after Sposito 1989)
Table 3.2.	Example of a K_d look-up table for uranium, uranium(VI), and uranium(IV)
Table 3.3.	Advantages, disadvantages, and assumptions of K_d determination methods and the assumptions in applying these K_d values to contaminant transport models
Table 5.1.	Chemical reaction models described in the literature 5.4
Table 5.2.	Examples of technical symposiums held on development, applications, and data needs for chemical reaction modeling
Table 5.3.	Component species in MINTEQA2 thermodynamic database 5.29
Table 5.4.	Organic ligands in MINTEQA2 thermodynamic database 5.31